Separation Science and Technology, 41: 123–147, 2006 Copyright © Taylor & Francis Group, LLC ISSN 0149-6395 print/1520-5754 online DOI: 10.1080/01496390500445626



Merrifield Resin Supported Chelate Forming Calix[4]arene-o-vanillinthiosemicarbazone Resin Employed for the Separation, Preconcentration and Trace Determination of Cr(VI), As(III) and Tl(I) in Water Samples

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Abstract: Calix[4]arene based polymer supported chelating resin **8** has been synthesized by covalently linking calix[4]arene-o-vanillinthiosemicarbazone via its 'lower rim' to Merrifield resin and was characterized by FT-IR, ¹H-NMR and elemental analysis. Resin **8** was competently used to separate and preconcentrate toxic metal ions Cr(VI), As(III) and Tl(I) in a column prior to their determination by UV-Visible and GF-AAS with R.S.D. between 1.0-1.4%. Various physico-chemical parameters like pH, concentration of eluting agents, flow rate, total sorption capacity, metal-ligand stoichiometry, exchange kinetics, preconcentration factor, distribution coefficient, breakthrough capacity, resin stability, effect of electrolytes, and associated metal ions have been studied. Detection limits ($3\sigma_B$) of 5.83, 8.48 and 5.56 µg L⁻¹, along with preconcentration factors of 117, 90 and 105 for Cr(VI), As(III) and Tl(I) respectively were achieved. Resin **8** was successfully applied to the

Received 17 August 2005, Accepted 10 October 2005

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separation and trace determination of Cr(VI), As(III) and Tl(I) from natural water samples of Ahmedabad city and the standard environmental, biological and geological reference materials.

Keywords: Calix[4]arene, o-vanillinthiosemicarbazone, preconcentration, Merrifield resin, GF-AAS

INTRODUCTION

The concentration of toxic metal ions like Cr(VI), Cu(II), Cd(II), As(III), Pb(II), Hg(II), Tl(I) etc. has increased beyond environmentally and ecologically sustainable levels due to the increase in industrial activity in the field of chemicals, dyes, color chemicals, petrochemicals, electroplating, textiles, and pharmaceuticals sectors over the years. It has resulted in severe contamination of ground and surface water; for e.g. rivers, lakes, rivulets etc. Especially in the developing countries, water and soil degradation generated by industrial effluents has been a serious issue (1). In order to achieve balanced growth, which is sensitive towards conservation of scarce natural resources, time-to-time analysis of these metal ions in natural water samples as well as in bio fluids is necessary. In natural water and biological systems Cr(VI), As(III), and Tl(I) are found in low levels and their separation and preconcentration are obligatory ahead of their trace quantification by instrumental or chemical methods of analysis. The metabolism and toxicology of Cr(VI), As(III), and Tl(I) are of great concern as they get accumulated in body systems. Their determination in food, blood, and tissues samples is also very important in order to identify cases of human intoxication leading to untold suffering in developing and least developed countries (2-4). Extraction of these metal ions is a tedious process as they are associated with a variety of complex species present in the natural aquatic systems. Many times they are found in such a low concentration that even sophisticated instrumental techniques like FAAS and ICP-OES are not sensitive enough to detect and quantify these toxic metal ions in natural water samples and their prior separation and preconcentration is required. Polymeric chelating ion exchange resins, which act as specific, sensitive and selective complexants for metal ions have been found to be very useful. For solid phase extraction of toxic metal ions, polymeric calixarenes (5-8) can play a vital role due to their characteristics such as lower toxicity, affordability, ease of synthesis, and operation. They are more eco-friendly compared to liquid-liquid extraction method where in disposal of organic waste has been a serious environmental hazard. Various calixarene derivatives have been recently studied with Tl(I) as these cavitands have been found useful in the complexation studies (9, 10). Thrust of the ongoing research programmes is on arsenic determination at trace and ultra trace levels and its speciation (11-21).

So far, there are no reports on the use of such polymeric calixarenes for separation and preconcentration of As(III), Cr(III), and Tl(I). Therefore, we planned to introduce o-vanillinthiosemicarbazone, a versatile chelating agent for toxic metal ions (22, 23) on the calix[4]arene scaffold, held by a polymer backbone (24). In this article, synthesis of Merrifield resin supported calix[4]arene-o-vanillinthiosemicarbazone (resin 8) and its application to the separation, preconcentration, and trace determination of Cr(VI), As(III) and Tl(I) in various natural water samples collected from Ahmedabad city and the standard environmental, biological and geological reference materials (SRMs) procured from National Institute of Standards and Technology (NIST, USA) have been described.

EXPERIMENTAL

Apparatus

Precoated silica gel plates (Merck 60 F_{254}) were used for TLC. FT-IR spectra were recorded on a Jasco infrared spectrophotometer as KBr pellets. Elemental analysis was done on Heraeus Carlo Erba 1108 elemental analyzer. ¹H-NMR spectra were recorded on Bruker operating at 100 MHz for proton in DMSO - d₆ with TMS as internal standard. A Systronics digital pH meter Model 335 was used for pH measurements. Flow rate in the column was adjusted with a Miclins peristaltic pump PP-10. Spectral measurements were done on a Hitachi 3210 UV-Visible spectrophotometer using 10 mm quartz cells. Atomic absorption measurements were performed on Perkin-Elmer model 420 atomic absorption spectrophotometer (AAS) pyrolytically coated HGA-76 graphite furnace (GF) for Cr(VI) and Tl(I). Total arsenic determination in the form of As(III) was done on a Chemito 201 AAS with a continuous flow hydride generator for arsine generation. Operating conditions for hydride generator were: argon flow rate-70 cm³/min; concentration of NaBH₄-1% w/v; concentration of HCl-5% v/v; sampling time-25 seconds; reaction time-25 seconds. Measurements were based on integrated absorbance.

	Cr(VI)	As(III)	Tl(I)
H.C. Lamp (mA)	10	10	10
Slit (nm)	0.5	0.5	0.5
Wavelength (nm)	357.9	193.7	276.8
Flame	Air-C ₂ H ₂	Air-C ₂ H ₂	Air-C ₂ H ₂
Fuel $(1 \min^{-1})$	2.0	2.0	2.0
Oxidant $(1 \min^{-1})$	8	10	8
Burner (cm)	10	10	10

Operating Conditions for AAS

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Reagents

All chemicals used in this work were of analytical grade of E. Merck or Lancaster. Merrifield peptide resin [chloromethylated polystyrene divinylbenzene co-polymer or CMPDB] (1% cross linked, 1.5 mmol equivalent of Cl/g of resin, 200–400 mesh) was procured from Aldrich. Glassware were soaked in 10% HNO₃ overnight before use and cleaned repeatedly with double distilled deionized water. Solvents were purified by the method of Weissberger et al (25). All aqueous solutions were prepared with quartz distilled deionized water, which was further purified by a Millipore Milli-Q water purification system. Standard stock solutions (2000 μ g ml⁻¹) of Cr(VI), As(III) and Tl(I) were prepared and standardized spectrophotometrically (26, 27). pH of solutions was maintained as described in the literature (28).

Synthesis (Fig. 1)

Synthesis of 5,11,17,23-tetratert-butyl-25,26,27,28-tetrahydroxycalix[4]arene 1 and25,26,27,28-tetrahydroxycalix[4]arene 2 were done as per the method of Gutsche et al. [48,49]. 26,28-dimethoxy-25,27-dihydroxycalix[4]arene 3 and 11,23-dinitro-26,28-di methoxy-25,27-dihydroxycalix[4]arene 4 were synthesized as reported by Arduini et al. [50].

Synthesis of Product 5 i.e.11,23-diamino-26,28-dimethoxy-25,27-dihydroxycalix[4]arene

Suspension of 5 g **4** (9.2 mmol), 10 ml hydrazine hydrate and a catalytic amount of Raney nickel in 100 ml methanol was refluxed for 6 hours in a 250 ml round bottomed flask. Reaction mixture was allowed to cool to room temperature, filtered and solvent was distilled under reduced pressure using a rotary evaporator. Residue was taken in 60 ml dichloromethane, washed thrice with 30 ml water and dried over MgSO₄. After filtration, solvent was distilled under reduced pressure using a rotary evaporator to give pure **5** in 3 g (68%) yield; mp > 300°C (dec.). FT-IR (KBr): 3180 cm⁻¹ (O-H stretching); 1252 cm⁻¹ (C-O stretching); 3435 cm⁻¹ (N-H stretching). ¹H-NMR (DMSO-d₆): 9.7 ppm (s,2,ArOH); 7.2 ppm (s,10,ArH); 3.4 ppm (br d,8,ArCH₂Ar); 3.7 ppm (s,6,OCH₃); 5.5 ppm (s,4,ArNH₂). Anal. calcd. for C₃₀H₃₀O₄N₂: C, 74.65; H, 6.27; N, 5.81. Found: C, 74.39; H, 6.08; N, 5.61.

Synthesis of Product 6 i.e.11,23-bis(chlorodiazenyl)-26,28-dimethoxy-25,27-dihydroxy calix[4]arene

10 g amine **5** (20.7 mmol) was diazotized in 35 ml HCl (40%) at 0°C in ice-bath with constant stirring for 30 minutes using 10 g NaNO₂ (145 mmol) in 100 ml H₂O:C₂H₅OH::1:1 media to yield **6**, which was then reacted further in situ.

Synthesis of Product 7 i.e. 11,23-bis(o-vanillinthiosemicarbazone)-26,28-dimethoxy-25,27-dihydroxycalix[4]arene

Solution containing **6** was taken in ice-bath in a beaker with an addition funnel supported over it containing 10 g o-VTSC (44.4 mmol) [Note 1] in 40 ml NaOH (15 g, 375 mmol) in water. Addition was completed in 1 hour at 0° C with constant stirring. Brownish-red colored product **7** was solidified by

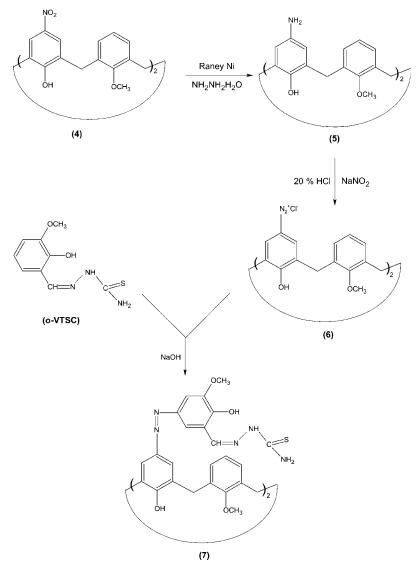


Figure 1. Reaction flowchart. (Continued)

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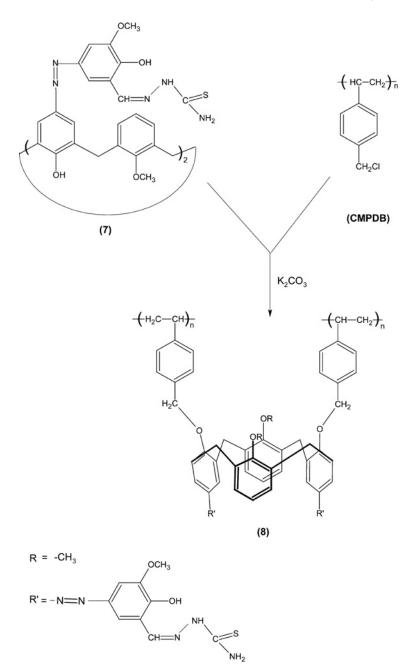




Figure 1. Continued.

evaporation of solvent. Purification was done by crystallization in iso-amyl acetate. Yield: 16 g (81%), mp > 300°C (dec.). FT-IR (KBr): 3464 & 3227 cm⁻¹ (O-H str.); 1591 cm⁻¹ (CH=N str.). ¹H-NMR (DMSO-d₆) (δ in ppm): 3.3 (s,8,ArCH₂Ar); 3.6–3.8 (s,12,OCH₃); 7.1–8.1 (s,14,ArH); 10.1 (s,4,ArOH). Anal. calcd. for C₄₈H₄₆O₈N₁₀S₂: C, 60.34; H, 4.86; N, 14.67; S, 6.72. Found: C, 59.96; H, 4.65; N, 14.76; S, 6.45.

[Note 1: Synthesis of 2-hydroxy-3-methoxybenzaldehyde thiosemicarbazone (o-VTSC) 5 g o-vanillin (33 mmol) was refluxed with 3 g thiosemicarbazide (33 mmol) in 80 ml, 95% ethanol at 75°C with constant stirring for 4 hrs. 0.5 g sodium acetate was added to maintain pH > 7. Reaction mixture was then cooled and filtered. Crude product was washed thrice with ethanol and water. Yield of pure o-VTSC was 7 g, mp 207°C.]

Synthesis of 8 by Loading of 7 on CMPDB

A mixture containing 10 g 7 (10.5 mmol), 10 g K₂CO₃ (72 mmol), and 10 g of CMPDB in 150 ml of DMF : THF::2 : 1 was heated at 75°C temperature for 8 hours under N₂ gas atmosphere. Thereafter polymeric beads of resin 8 were separated through a porous filter at pump. Repeated washing with DMF and THF followed by water were carried out to complete remove unreacted 7 and weight of dried resin 8 was found to be 12.34 g. Amount of 7 loaded on CMPDB was evaluated by difference of mass taken before and after coupling for dried resin and was found to be 23.4%

Percentage of 7 hooked on C	MPD	B = -	$\frac{W_2 - W_1}{W_1} \times 100\%$
where, weight of dried CMP	DB	:	(W ₁)
weight of dried resin	8	:	(W ₂)

Elemental analysis of nitrogen and sulphur in resin **8** were found to be 2.78% and 1.26% respectively, which further confirmed amount of loading to be 23-24%. It shows that only 0.4 mmol i.e. 26.5% of total available Cl (1.5 mmol/g of CMPDB) reacted with 0.2 mmol **7**.

Procedure for Column Concentration and Determination of Cr(VI), As(III) and Tl(I)

A glass column ($200 \times 8.0 \text{ mm}$) was prepared by adding slurry of beads of same mesh size having 1 g resin **8** in CH₃OH : H₂O::1 : 1. Height of column section containing resin **8** (bed height) was 15 mm. It was conditioned with different buffer solutions for samples containing Cr(VI), As(III) and Tl(I) prior to their elution from column. Suitable aliquot of solution containing 0.01-3.03, 0.065-7.2, and $0.023-4.9 \,\mu g \, ml^{-1}$ of Cr(VI), As(III) and Tl(I) respectively was passed through column after adjusting appropriate pH at an optimum flow rate. Stripping of metal ions from resin bed was carried

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			Metal lons	
No.	Parameters	Cr(VI)	As(III)	Tl(I)
1	pH Range	3.0-4.0	4.5-5.5	5.5-6.5
2	Flow rate $(ml min^{-1})$	1.5	1.0	1.0
3	Concentration of acid for desorption	0.1 N H ₂ SO ₄	0.5 N HCl	1 N HNO ₃ / 3 N HCl
4	Total Sorption capacity $(\mu g g^{-1} \text{ of Resin})$	11325	15575	34397
5	Distribution coefficient (K _d)	3360	2600	2128
6	Preconcentration factor (PF)	117	90	105
7	Breakthrough capacity $(\mu g g^{-1} \text{ of resin})$	2151	2647	5503
8	R.S.D. $(\%)^a$	1.2 - 1.4	1.1-1.3	1.0 - 1.2
9	Average recovery (%)	98	99	97-98
10	$t_{1/2}$ for exchange (minutes)	6.5	9.0	12.5

Table 1. Parameters optimized for sorption and desorption of Cr(VI), As(III) and TI(I) on resin 8

^aTen determinations.

out by suitable mineral acids. Eluants were collected in a 25 ml volumetric flask, made to volume by double distilled water and determined spectrophotometrically as well as by GF-AAS (Table 1).

RESULTS AND DISCUSSION

Parameters Optimized for Separation and Preconcentration of Cr(VI), As(III) and Tl(I)

1. Effect of pH

A glass column was packed with 1 g resin **8**. 100 ml solutions containing 3, 7, and 4.5 μ g ml⁻¹ Cr(VI), As(III) and Tl(I) respectively were passed through the column at their optimum flow rate by varying pH. Metal sorption (%) was measured by desorbing resin **8** with a suitable eluting agent and then determined by spectrophotometry and GF-AAS. Optimum pH for Cr(VI), As(III) and Tl(I) were found to be 3.5, 5.0, and 6.0 respectively (Table 1, Fig. 2). Their selectivity in pH of sorption suggests possibility of separation of these metal ions in presence of each other in column.

 Effect of Flow Rate Effect of flow rate on extent of metal ion sorbed on resin was studied in column packed with 1 g resin 8 at 30°C. Feed solutions containing 3, 7,

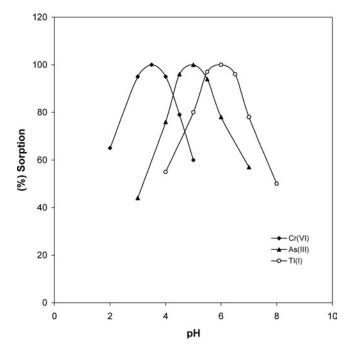


Figure 2. Effect of pH on Sorption of Cr(VI), As(III) and TI(I) by resin **8**. Amount of resin **8** in column: 1 g; Volume of metal ion solution passed: 100 ml; Cr(VI): $3 \mu g m l^{-1}$, Elution by: 0.1 N H₂SO₄; As(III): $7 \mu g m l^{-1}$; Elution by: 0.5 N HCI; Tl(I): 4.5 $\mu g m l^{-1}$, Elution by: 1 N HNO₃.

and 4.5 μg ml⁻¹ Cr(VI), As(III) and Tl(I) respectively were passed at different flow rates (0.5, 1.0, 1.5, 2.0 etc. ml min⁻¹) maintained by peristaltic pump. Flow rate of liquid during chromatographic separation should be such that working conditions are close to equilibrium state as flow rate depends upon particle size and temperature. Optimum flow rate may be defined as rate of flow of effluent through column at which more than 98% sorption takes place. Optimum flow rates obtained were 1.5, 1.0, and 1.0 ml min⁻¹ for Cr(VI), As(III) and Tl(I) respectively. An increase in flow rate resulted in decrease in sorption due to insufficient contact time between resin phase and metal ion solution (Table 1, Fig. 3).
Effect of Concentration of Eluting Agents

Effect of various eluting agents like HCl, HNO₃ and H₂SO₄ was studied by taking their different concentrations. 1 g resin **8** in column was conditioned at pH of maximum sorption and then fed with 100 ml solutions containing 3, 7 and 4.5 μ g ml⁻¹ Cr(VI), As(III) and Tl(I) respectively. Metal ions were desorbed with different mineral acids and then determined by spectrophotometry and GF-AAS (Table 1 & 2). It was observed that quantitative elution was possible with 15 ml of 0.1 N

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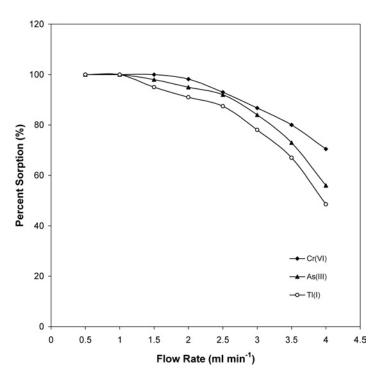


Figure 3. Effect of Flow Rate on Sorption of Cr(VI), As(III) and TI(I) by resin 8. Amount of resin 8 in column: 1 g; Cr(VI): $3 \mu g m l^{-1}$; pH 3.5; As(III): $7 \mu g m l^{-1}$; pH 5.0; Tl(I): $4.5 \mu g m l^{-1}$; pH 6.0.

 H_2SO_4 , 0.5 N HCl, and 1 N HNO₃/3 N HCl for Cr(VI), As(III), and Tl(I) respectively.

4. Total Sorption Capacity

Total sorption capacity studies were carried out using batch techniques for Cr(VI), As(III) and Tl(I). 1 g resin **8** was equilibrated with 500 ml solution containing 100 μ g ml⁻¹ metal ions at their optimum pH for 2 hours at 30°C. Amount of metal ions sorbed on resin **8** was calculated from difference in metal ion concentration before and after desorption (Table 1). Total sorption capacities of resin were found to be 11325, 15575, and 34397 μ g g⁻¹ for Cr(VI), As(III) and Tl(I) respectively.

5. Equilibrium Distribution Coefficient K_d Exchange equilibria are very often expressed in terms of distribution coefficient K_d. This quantity is given by ratio of equilibrium concentrations of same metal ion in resin phase and in solution. K_d of metal ions between resin 8 and aqueous phase was determined by batch experiment. 1 g resin 8 was equilibrated with 25 ml solution containing not more than 453, 623 and 1376 µg ml⁻¹ Cr(VI), As(III) and Tl(I) respectively for 2 hours at 30°C. Solution was filtered to remove resin 8 and filtrate was

	_	HCl		_	HNO ₃			H_2SO_4	
Conc. (N)	Cr(VI) (%)	As(III) (%)	Tl(I) (%)	Cr(VI) (%)	As(III) (%)	Tl(I) (%)	Cr(VI) (%)	As(III) (%)	Tl(I) (%)
0.01	5.5	22.8	5.0	5.3	4.3	24.5	66.6	4.8	6.0
0.1	12.5	66.5	9.7	11.5	16.3	48.8	99.8	11.1	9.0
0.5	17.1	99.7	12.0	27.7	19.3	69.0	_	16.8	13.1
1.0	21.6		17.9	32.5	21.6	99.9	_	22.7	18.2
1.5	29.5		32.3	44.6	27.5		_	30.5	28.5
2.0	40.7		63.0	50.1	36.3		_	38.2	39.1
3.0	49.9		99.8	58.1	42.8		_	45.9	52.3

Table 2. Effect of concentration of eluting agents for desorption of Cr(VI), As(III) and Tl(I) from resin **8**

^{*a*}Average of ten determinations.

Experimental condition: Resin: 1 g; Volume of solution passed: 100 ml; Cr(VI): pH 3.5; $3 \mu g m l^{-1}$; As(III): pH 5.0; $7 \mu g m l^{-1}$; Tl(I): pH 6.0; $4.5 \mu g m l^{-1}$.

subjected to spectrophotometric and GF-AAS determination of metal ion content. Concentration of eluted metal ion solution was also measured (Table 1). K_d for Cr(VI), As(III) and Tl(I) were found to be 3360, 2600 and 2128 respectively. K_d or efficiency of resin **8** was measured using equation,

 $K_{d} = \frac{\text{Amount of metal ion taken up by resin}}{\text{Amount of metal ion remaining in solution}} \\ \times \frac{\text{Volume of solution}}{\text{Amount of resin taken}}$

Plots of log K_d against pH (Fig. 4) were straight lines with slopes nearly equal to unity, indicating 1:1 stoichiometry between metal ions and resin **8** (29).

6. Exchange Kinetics

Batch experiments were carried out to determine rate of loading of Cr(VI), As(III) and Tl(I) on resin **8**. 1 g resin **8** was stirred with 100 ml solution containing 113, 156 and 344 μ g ml⁻¹ Cr(VI), As(III) and Tl(I) respectively at 30° C. 5 ml of solution was withdrawn at predetermined intervals and analyzed. Concentration of metal ion in aqueous phase was determined by spectrophotometry and GF-AAS. Amount of metal ion loaded on resin phase was calculated by measuring concentration of 50% of metal ions (t_{1/2}) was found to be 6.5, 9.0 and 12.5 minutes for Cr(VI), As(III) and Tl(I) respectively, which indicated very good accessibility of these metal ions towards chelating sites.

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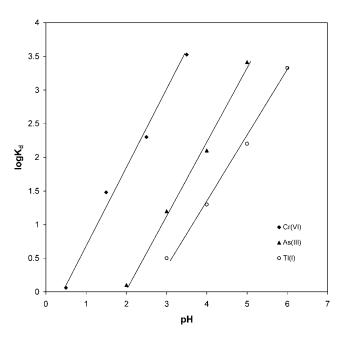


Figure 4. Plot of $\log K_d$ against pH for Sorption of Cr(VI), As(III) and TI(I) on Resin 8. Amount of resin 8: 1 g; Volume of solution: 25 ml; Cr(VI): 453 µg ml⁻¹; As(III): 623 µg ml⁻¹; Tl(I): 1376 µg ml⁻¹.

7. Breakthrough Studies

Breakthrough capacities are more significant and useful than total sorption capacities in ion exchange chromatographic applications as it gives actual working capacity of resin in column. Breakthrough capacity or effective capacity may be defined as capacity at moment when adsorbed component has appeared in effluent with an increased concentration. It is one of most critical parameters when working under dynamic condition. Breakthrough studies were carried out by taking 1 g resin 8 in column having 200 mm length and 8.0 mm internal diameter with bed height of 15 mm and passing $10 \,\mu g \,\mathrm{ml}^{-1}$ of metal ions [Cr(VI), As(III) or Tl(I)] at their optimum pH and flow rate. An aliquot of 2.0 ml eluant was collected each time and analyzed by spectrophotometry and GF-AAS for metal ion content (Fig. 6). Breakthrough capacities of resin 8 for Cr(VI), As(III) and Tl(I) were found to be 2151, 2647 and 5503 μ g g⁻¹ respectively (Table 1). All curves were steep at their breakthrough points at different times, indicating possibility of separating these metal ions from their mixtures.

 Preconcentration of Cr(VI), As(III) and Tl(I) Concentration of trace elements in natural water is too low for their direct determination. Therefore, preconcentration or enrichment step is

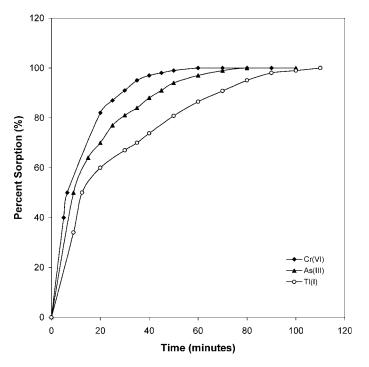


Figure 5. Exchange kinetics of Cr(VI), As(III) and TI(I) on resin **8**. Amount of resin **8**: 1 g; Volume of feed solution: 100 ml; Cr(VI): 113 μ g ml⁻¹; pH 3.5; As(III): 156 μ g ml⁻¹; pH 5.0; Tl(I): 344 μ g ml⁻¹; pH 6.0.

unavoidable even with a sensitive detection method. Resin **8** was studied for column concentration of Cr(VI), As(III) and Tl(I) in terms of their preconcentration factor (PF).

$PF = \frac{Concentration of metal in stripping solution}{Initial concentration of metal in feed solution}$

1000 ml solutions containing 6, 9, 6 μ g l⁻¹ Cr(VI), As(III) and Tl(I) at pH 3.5, 5.0, 6.0 respectively, were passed through column containing 1 g resin **8**. Preconcentrating ability of resin **8** was assessed from elution profile of metal ions by plotting concentration of effluents as a function of volume of stripping solutions: 8.5 ml, 0.1 N H₂SO₄ for Cr(VI); 11.1 ml, 0.5 N HCl for As(III); 9.5 ml, 1 N HNO₃ for Tl(I) respectively. Preconcentration factors were found to be 117, 90 and 105 for Cr(VI), As(III) and Tl(I) respectively with 98-99% recovery (Table 1 & 3, Fig. 7).

9. Resin 8 Stability Tests

Stability of resin 8 was tested to evaluate its sustainability for repeated cycles of sorption and desorption with different mineral acids. It was

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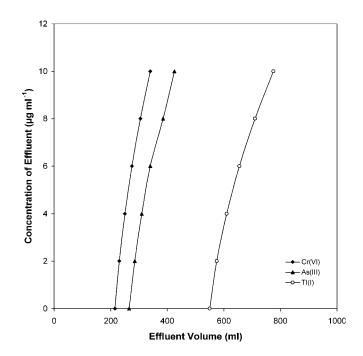


Figure 6. Breakthrough curves for Cr(VI), As(III) and TI(I) on resin **8**. Amount of resin **8** in column: 1 g; Concentration of solution passed: $10 \,\mu g \,ml^{-1}$; Cr(VI): pH 3.5; As(III): pH 5.0; Tl(I): pH 6.0.

observed that there was no detectable change in physical and chemical properties of resin **8** in concentrated acidic solutions of 5.0 N HCl, 3.0 N HNO₃, and 3.5 N H₂SO₄. To check regenerating capacity of resin **8**, it was subjected to repeated sorption and desorption of Cr(VI), As(III), and Tl(I) at their optimum conditions in column. 1 g resin **8**

Metal ions	Volume of solution passed (ml)	Concentration of solution passed (µg1 ⁻¹)	Volume of eluted solution (ml)	Recovery (%)	Preconcentration factor (PF)
Cr(VI)	1000	6	8.5	98.0	117
As(III)	1000	9	11.1	99.0	90
Tl(I)	1000	6	9.5	98.0	105

Table 3. Preconcentration factors for the Sorption of Cr(VI), As(III) and Tl(I) on resin 8

^{*a*}Average of ten determinations.

Experimental Condition: Resin: 1 g; Cr(VI): pH 3.5; Elution by 0.1 N H₂SO₄; As(III): pH 5.0; Elution by 0.5 N HCl; Tl(I): pH 6.0; Elution by 1 N HNO₃.

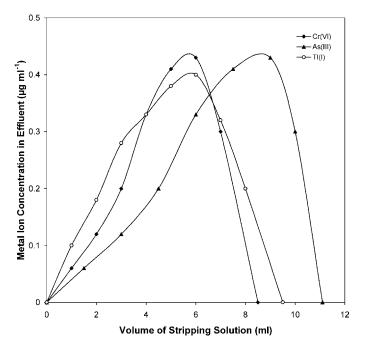


Figure 7. Elution Profile of Cr(VI), As(III) and TI(I) on resin **8**. Amount of resin **8** in column: 1 g; Volume of solution passed: 1000 ml; Cr(VI): $6 \mu g l^{-1}$; pH 3.5; elution by 0.1 N H₂SO₄; As(III): $9 \mu g l^{-1}$; pH 5.0; elution by 0.5 N HCI; Tl(I): $6 \mu g l^{-1}$; pH 6.0; elution by 1 N HNO₃.

was taken in column and 500 ml solution containing 100 μ g ml⁻¹ metal ions was passed through it at 30°C. Resin **8** was then stripped off metal ions with 25 ml of suitable eluant. Resin **8** showed very good stability up to 15 repeated cycles of sorption and desorption (Fig. 8), after which it showed 2–3% decrease in efficiency, which may be due to hydrolysis of resin **8** when treated with mineral acids.

- 10. Effect of Electrolytes and Associated Cations
 - Limit of tolerance of anions and cations on the sorption of Cr(VI), As(III) and Tl(I) is defined as that concentration which causes an error of 2-3% in the recovery of these metal ions. Effect of commonly associated cations in natural water samples such as Fe(II), Fe(III), Co(II), Ni(II), Mn(II), alkali [Li(I), Na(I), K(I), Rb(I), Cs(I)], and alkaline earth [Mg(II), Ca(II), Sr(II), and Ba(II)] had been studied under optimum conditions of metal ion sorption. Mn(II) did not affect sorption of Cr(VI), As(III) and Tl(I) up to 4 mg, however, for other cations tolerance limits were much higher (Table 4). Effect of anions and their limit of tolerance on sorption of Cr(VI), As(III) and Tl(I) by resin **8** was studied by taking different concentrations of some electrolytes like NaCl, NaF, NaBr, NaNO₂, NaNO₃, Na₂SO₄, Na₃PO₄, and CH₃COONa. Results (Table 5) showed that

138 V. K. Jain et al. 40000 35000 -Cr(VI) 5000 -As(III) 0 2 0 4 6 8 10 12 14 16 **Cycle Number**

Figure 8. Evaluation of stability of resin 8 for Cr(VI), AS(III) and TI(I) by Sorption and elution. Amount of resin **8** in column: 1 g; Volume of feed solution: 500 ml; Concentration of feed solution: $100 \,\mu g \, ml^{-1}$; Cr(VI): pH 3.5; elution by 0.1 N H₂SO₄; As(III): pH 5.0; elution by 0.5 N HCI; Tl(I): pH 6.0; elution by 1 N HNO₃.

			Ι	nterfering	g ions (mg	g)	
Metal ions (100 µg each)	Fe(II)	Fe(III)	Co(II)	Ni(II)	Mn(II)	Alkali metal ions	Alkaline earth metal ions
Cr(VI) As(III) Tl(I)	16 14 7	21 17 15	17 16 11	20 15 14	4 4 5	DNI DNI DNI	DNI DNI DNI

Table 4. Effect of commonly associated cations in water samples on the Sorption of Cr(VI), As(III) and Tl(I) on resin 8

^aAverage of ten determinations. DNI: Do not interfere.

Experimental Condition: Resin: 1 g; Volume of solution passed: 100 ml; Cr(VI): pH 3.5; As(III): pH 5.0; Tl(I): pH 6.0.

			Cor	ncentratio	n of elec	trolytes (r	$\operatorname{nol} l^{-1}$)	
Metal ions $(2.5 \mu g \mathrm{ml}^{-1})$	NaF	NaCl	NaBr	NaNO ₂	NaNO ₃	Na ₂ SO ₄	Na ₃ PO ₄	CH ₃ COONa
Cr(VI)	1.5	2.3	2.5	3.0	2.5	1.3	0.2	2.5
As(III)	2.0	3.0	2.0	2.1	2.8	1.2	0.3	2.3
Tl(I)	1.2	2.0	1.4	2.2	3.0	1.7	0.4	2.1

Table 5. Tolerance limits of electrolytes on the Sorption of Cr(VI), As(III) and Tl(I) on resin 8

^aAverage of ten determinations.

Experimental condition: Resin: 1 g; Volume of solution passed: 100 ml; Cr(VI): pH 3.5; As(III): pH 5.0; Tl(I): pH 6.0.

except Na_3PO_4 , all other electrolytes did not interfere between 1.0-3.0 M concentration range, which further augments potential application of resin **8** for analysis of real samples.

Separation of Cr(VI), As(III) and Tl(I)

Resin 8 was found useful for separating chromium, arsenic and thallium from mixture due to favourable kinetics and different breakthrough capacities in column. Resin 8 was tested for simultaneous separation of their ternary mixture. Hence, their mixture (each 100 µg in 25 ml buffer solution) was passed through column at optimum pH and flow rate. Column effluents were analyzed for metal ions by spectrophotometry and GF-AAS. In separation of their ternary mixture containing 100 µg each of Cr(VI), As(III), and Tl(I) was passed through column maintained at pH 3.0 at flow rate of 1.0 ml min⁻¹. At this pH, most of As(III) and Tl(I) came out with effluent while Cr(VI) was retained in column. Cr(VI) was eluted with 0.1 N H₂SO₄. Small amounts of As(III) and Tl(I) ($\leq 2\%$) which were sorbed by resin 8 were then separated on the basis of selective elution with different eluting agents. As(III) was eluted first with 0.5 N HCl followed by Tl(I) with 1 N HNO₃. Effluent containing As(III) and Tl(I) was then passed through a fresh column maintained at pH 5.5. Both metal ions were retained in column and As(III) was desorbed first with 0.5 N HCl followed by Tl(I) with 1 N HNO₃. Thus, quantitative separation was achieved as seen in separation pattern (Fig. 9).

Analytical Performance (Calibration Graph, Detection Limit and Precision)

Standard stock solutions of these elements were used after appropriate dilution for preparation of calibration standards for Cr(VI), As(III), and Tl(I). Precision of proposed method was examined based on optimum

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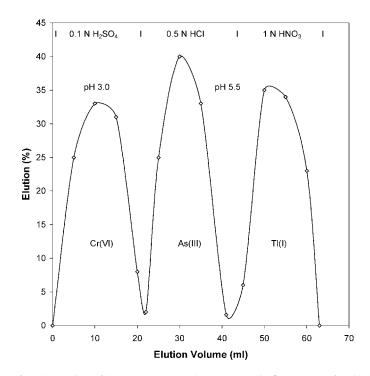


Figure 9. Separation of Cr(VI), As(III) and TI(I) on Resin 8. Amount of resin 8: 1 g; Column maintained first at pH 3.0 and than at pH 5.5; Cr(VI): 100 μ g in 25 ml buffer; As(III): 100 μ g in 25 ml buffer; TI(I): 100 μ g in 25 ml buffer.

condition for preconcentration in column and expressed as R.S.D.; 1.3% for Cr(VI), 1.2% for As(III), and 1.1% for Tl(I). Limit of detection (LOD) was defined as that analyte concentration giving a signal equal to three times standard deviation of blank signal (42). Limit of quantification (LOQ) was then set as ten times standard deviation of blank signal (Table 6).

Table 6. Regression analysis for the calibration graph, linear range, fetection limit and precision

Metal Ions	Linear range $(\mu g m l^{-1})$	Regression equation	Correlation coefficient	$\begin{array}{c} LOD \\ (\mu g l^{-1}) \end{array}$	$\begin{array}{c} LOQ \\ (\mu g l^{-1}) \end{array}$	R.S.D. ^a (%)
Cr(VI)	0.010-3.03	$A = C \times 0.599$ - 0.0011	1.0010	5.83	19.43	1.3
As(III)	0.065-7.20	$A = C \times 0.201$ - 0.0076	0.9979	8.48	28.26	1.2
Tl(I)	0.023-4.90	$A = C \times 0.3469$ - 0.0006	1.0110	5.56	18.56	1.1

^{*a*}Ten determinations. A = Absorbance, C = Concentration ($\mu g m l^{-1}$).

		Cr(VI)	(1/	As(III)	(III)	T1(I)	[]
Sample	Method	Amount $(\mu g l^{-1})$	$R.S.D.^{a}$ (%)	Amount $(\mu g l^{-1})$	$R.S.D.^{a}$ (%)	Amount $(\mu g l^{-1})$	$R.S.D.^{a}$ (%)
Sabarmati River, Near	Present method	42	1.0	39	1.1	25	1.1
Thermal Power	Standard addition	43	1.1	38	1.2	24	1.3
Station, Ahmedabad							
Ground Water,	Present method	20	1.4	29	1.0	19	1.3
University Area,	Standard addition	19	1.2	30	1.2	20	1.1
Ahmedabad							
Ground Water, Vatva	Present method	94	1.2	67	1.0	40	1.2
Industrial Zone,	Standard addition	96	1.3	69	1.1	39	1.2
Ahmedabad							

Table 7. Determination of Cr(VI), As(III) and Tl(I) in natural water samples on resin 8

^aTen determinations.

Experimental condition: Resin: 1 g; Sample volume: 1000 ml.

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	Metal ion cond	centration ^a found b	y GF-AAS
Sample	$\frac{Cr(VI)}{(\mu g g^{-1})}$	$As(III) (\mu g g^{-1})$	$Tl(I) \\ (\mu g g^{-1})$
SRM 361	6937.5 (6940)	168.9 (170)	
(Low alloy steel)			
SRM 2711	47.2 (47)	104.1(105)	2.39 (2.47)
(Montana soil)			
SRM 1573a	1.99 (1.99)	0.111 (0.112)	
(Tomato leaves)			
SRM 1577b	_	0.049 (0.047)	
(Bovine lever)			

Table 8. Determination of Cr(VI), As(III) and Tl(I) in Standard reference materials (SRM) of NIST on resin **8**

(Certified Values in the Brackets)

^aThree determinations.

Experimental condition: Resin: 1 g; Sample: 1 g.

Application to Analysis of Water Samples

Validity of system was established by subjecting resin $\mathbf{8}$ to natural water samples and SRMs procured from NIST. Results obtained for water samples by proposed method were compared with that of standard addition method. Natural water samples and aqueous samples of reference standards were prepared as per methods described in experimental section and then passed through column as explained in the general procedure for column concentration. SRMs were decomposed in mixture of acids as described in literature (30). Amount of metal ions present in water samples and their R.S.D. values are given in Table 7 and method validation with respect to NIST standards has been given in Table 8. Higher efficiency of separation curves with lower R.S.D., suggests utility of resin $\mathbf{8}$ for various analytical applications in determination of Cr(VI), As(III) and Tl(I).

CONCLUSION

Polymer supported calixarenes have established themselves with prospective ionophoric properties that can be used for varied analytical applications. Resin **8** appears capable for trace enrichment of metal ions as noticeable from column studies. Separation curves exhibit efficiency of resin **8** for real-time separation of Cr(VI), As(III) and Tl(I) in presence of each other with minimum interference from alkali and alkaline earth metal ions. Resin **8** shows high separating and preconcentration ability for these metal ions compared to other preconcentration methods (31-45) reported in literature as summarized in Table 9. Resin **8** can be successfully regenerated for 15

Res	in Support	ed Ch	elate i	in Wate	er \$	Samples	•					
	References	(31)	(32) (33)	(34)		(35)	(36)	(37)	(38)	(39)		(40)
	Application	Tap water and biological samples	Fresh water samples Water samples	Ground water samples		Sea water and Reference material (NASS-4)	Environmental water samples	Potable water samples	Parental solutions	Water samples		Water samples
d thallium	Detection limit	0.01 ppm	0.1 ppm 0.02 ppm	30.6 ppm		0.008 ppm	0.244 & 0.235 ppm	0.027 ppb	0.029 ppb	1.9 &	2.3 ppm	7.4 ppb
, chromium, and	Precon- centration factor	52.5		l		44		25	70	20.8 <i>&</i>	24.9	I
reported for arsenic	Analytical technique	ET-AAS	ICP-AES ICP-AES	Potentiometric flow	analysis	ET-AAS	Derivative F-AAS	ET-AAS	ICP-OES	F-AAS		F-AAS
Table 9. Summary of preconcentration methods reported for arsenic, chromium, and thallium	Preconcentration procedure	Cloud Point Extraction (Triton X-114)	Anion Exchange Resin Co-precipitation	Solid Adsorbent		PTFE knotted reactor	On-line flow injection	Dowex-1 Oxalate form	On-line flow injection	Zirconium(IV) Phos-	phate Immobilized on Silica gel	Saccharomyces Carls- bergensis immobilized on amberlite XAD-4
Table 9. Summary	Metal ions	As(III) & As(V)	As(III) & As(V) As(III) & As(V)	As(V)		As(III)	Cr(III) & Cr(VI)	Cr(VI)	Cr(III) & Cr(VI)	Cr(III) & Cr(VI)		Cr(III)

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(continued)

Table 9. Continued

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Metal ions	Preconcentration procedure	Analytical technique	Precon- centration factor	Detection limit	Application	References
	Bentonite modified with trioctvlamine	F-AAS		1.7 ppm	Waste water samples	(41)
	Agrobacterium tumefa- cients immobilized on amberlite XAD-4	F-AAS	25	3.6 ppb	Water samples, alloys, infant foods and stan- dard reference materials	(42)
	Nanometer-sized alumina	ICP-AES	5		Environmental reference materials	(43)
	Polyurethane foam	Platform-GF- AAS		0.05 ppm	Urine	(44)
		ICP-MS	20		Sea water samples	(45)
Cr(III), As(III) &	Merrifield resin sup-	GF-AAS	90,117 &	8.48, 5.83	Natural water samples $\&$	Present
	ported calix[4]arene		105	&	environmental, bio-	method
	derivative			5.56 ppb	logical & geological reference materials	

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cycles of operation in column. Results obtained from analysis of natural water samples and SRMs prove reliability of proposed method for its application to a range of samples for their determination by GF-AAS.

ACKNOWLEDGEMENT

Authors are thankful to GUJCOST, Gujarat for financial assistance. Authors also acknowledge CDRI, Lucknow for instrumental facilities and INFLIBNET, Ahmedabad for e-journals.

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